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Research Article Effects of Polyoxyethylene Sorbitan Monooleate on Degradation of Pyrene in Soils Growing Sorghum Sudanese

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Abstract: Phytoremediation is becoming a cost-effective technology for the *in-situ* clean up of sites polluted with Hydrophobic Organic Contaminants (HOCs). The major factors limiting phytoremediation are the mass transfer, rate of plant uptake and microbial biodegradation of HOCs. To evaluate the potential of surfactants to enhance phytoremediation for HOC-contaminated sites, the efficacy of Sudan grass, at the absence or presence of polyoxyethylene sorbitan monooleate (Tween 80), on the degradation of pyrene in soils were investigated and mechanisms of Surfactant-Enhanced Phytoremediation (SEPR) were discussed. Results showed that the presence of Tween 80 enhanced dissipation of pyrene at initial contents ranging from 20.24 to 321.42 mg/kg. During the 70-d SEPR-experiments, about 801.84~539.99‰ of pyrene was removed from planted soils, only 242.28~122.79‰ degradation of pyrene occurred in unplanted ones. With the presence of Tween 80, the dissipation ratios of pyrene in planted ones were increased up to 863.94~609.63‰, which was 77.27~129.14‰ higher than those in corresponding soils without surfactant. Among all possible pathways, contribution of plant-microbial interactions on dissipation of pyrene was the most significant, either at the presence (456.73‰) or absence (515.58‰) of Tween 80, were the primary means of contaminant degradation. Results suggested SEPR may be a feasible way for reinforcing removal of HOCs in contaminated sites.

Keywords: Hydrophobic organic contaminants, pyrene, phytoremediation, surfactants, SEPR, tween 80

INTRODUCTION

Phytoremediation is a promising approach to soil remediation due to its convenience, cost-effectiveness and environmental acceptability (Yi and Crowley, 2007). Plants may contribute to the dissipation of organic contaminants through an increase in the number of microbes, improvement of soil properties and structure, promotion of humification and adsorption of pollutants in the rhizosphere, but the impact of each process has not been clearly elucidated and remediation efficacy varies greatly among plant species depending differences the of soil conditions on and physicochemical nature of contaminants (Siciliano et al., 2003). However, there also exist many limitations for large-scale application of this technology. One serious limitation is that HOCs in soils usually exhibit low bioavailability to both microorganisms and plants due to their strong affinity to the soil matrix, especially to soil organic matter, which would limit the application of phytoremediation (Chen et al., 2005; Eriksen-Hamel and Whalen, 2008).

Surfactants are amphiphilic molecules with a hydrophobic portion (usually a long hydrocarbon chain) and a hydrophilic (polar) head group. Surfactant is an abbreviation for 'surface active agent,' so named because these molecules tend to migrate to surfaces and interfaces or create new molecular surfaces by forming aggregates. The Critical Micelle Concentration (CMC), the surface and Interfacial Tension (IFT) and the Hydrophile-Lipophile Balance (HLB) are the three main parameters that help characterize surfactant activity in solutions. There are four general classes of surfactants, which are classified by the charge on the head group: anionic, nonionic, cationic, polar zwitterionic (Lee et al., 2002). Surfactants may aid in remediation by decreasing the surface tension of the water and efficiently solubilizing or otherwise mobilizing nonaqueous phase liquids. As the surfactant concentration increases, the surface tension of the solution decreases until the CMC is reached. The addition of surfactants to a soil contaminated with HOCs might increase the mass transfer and the bioavailability of these compounds, facilitating their degradation. The ability of the surfactants to increase

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desorption, apparent aqueous solubility and microbial bioavailability of HOCs has been well documented. Recently, it has been shown that addition of surfactants facilitates the uptake of HOCs by plants. These results indicate a promising opportunity to use surfactants to enhance phytoremediation efficacy.

In this study, effect of Sudan grass (Sorghum sudanense), at the absence or presence of polyoxyethylene sorbitan monooleate (Tween 80), on the dissipation of pyrene (Pyr) in soils was investigated and each removal pathway of Pyr in the process of remediation was compared in order to understand the potential of Surfactant-Enhanced Phytoremediation (SEPR) of HOC-contaminated soils. Furthermore, the mechanisms of SEPR for soil HOCs (including HOC desorption, microbial biodegradation and plant uptake) were assessed by considering experimental observations.

MATERIALS AND METHODS

Chemicals: Pyrene (Pyr), as representative of HOCs, was obtained from Sichuan University with a purity >98.5%, molecular weights of 202.26 g/mol and 4.88 of $\log K_{ow}$ (K_{ow} , octanol-water partition coefficient).

Surfactants: Polyoxyethylene sorbitan monooleate (Tween 80), as representative of nonionic surfactants, was purchased from Tokyo Kasei Kogyo Co. Ltd with a purity≥98%, molecular weights of 1310 g/mol, 15 of HLB and CMC of 15.72 mg/L. Four percent (vol/vol) was selected in this study because it showed a good solubilization for hydrophobic organic compounds and gave the best removal efficiency in a previous study.

Contaminated soils: Soils tested were collected from 6~18 cm horizontal with pH 6.78, 1.46% organic matters and originally free of PAHs. After being airdried and sieved through a 2 mm mesh, soils were spiked with Pyr in acetone. When acetone evaporated off, spiked soils were progressively mixed with unpolluted soils and homogenized. The initial Pyr in soils were measured and followed as: C₀ (free PAHs), C₁ (20.24±0.94 mg/kg), C₂ (39.58±1.51 mg/kg), C₃ (79.86±2.37 mg/kg), C₄ (160.64 ± 3.05 mg/kg) and C₅ (321.42±4.93 mg/kg). Treated soils were packed into pots (2.0 kg) and equilibrated 4d under 45% of the water holding capacity (WHC).

Plant: After surface sterilized in 10% H₂O₂ for 10 min and rinsed with sterile distilled water, Sudan grass seeds were germinated and grown in moist perlite and seedlings of uniform size were transplanted to the designated greenhouse pots 7d after emergence.

Experimental design: Experiments were conducted in an intelligent greenhouse. They were divided into group A and group B according to studied objects. In group A, four different treatments with five replicates were performed:

- Treatments 1 (CK₁), unplanted microbe-inhibited soils, where 0.1% NaN₃ was added to inhibit the microbial activity (Wei and Pan, 2010)
- Treatments 2 (CK₂), unplanted soils
- Treatments 3 (TR₃), planted microbe-inhibited soils added 0.1% NaN₃
- Treatments 4 (TR₄), planted soils. Seedlings were thinned 5 d after transplanting to eight plants per pot. WHC was checked and adjusted regularly with sterilized water to 50% (Wei and Pan, 2010).

In Group B, the same established treatments and culture conditions were used. Based on soil properties and weights per pot, 300 ml of 4% (vol/vol) aqueous surfactant solution (Tween 80) were added to soils at the beginning of experiments and then equilibrated 4d under 45% WHC prior to transplanting.

Soils and plants were destructively sampled after the 70d SEPR-experiments. Shoot and root tissues, separated from soil, were washed with distilled water and dried with filter paper. The plant materials and soil samples were stored at -40.0° C to prevent microbial degradation of contaminants. An aliquot of each soil or plant sample was weighed, dried at 105.0°C for 24 h and weighed again.

Determination of HOCs: The HOCs contents were determined using the high performance liquid chromatography method with ultraviolet detection after a preliminary sample treatment with ultrasonic techniques and the detailed methods to extract Pyr have been described (Pan *et al.*, 2014).

Prior to use, all methods were tested for efficiency of recovery. For PAHs-amended soils, plant tissue samples, recovery averaged 93.52 % (n = 9, RSD<6.37%; RSD, relative standard deviation) and 93.23% (n = 9, RSD< 5.49%), respectively.

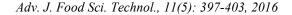
Data analyses: The obtained data were analyzed using SPSS version 12.0 and levels of significance were assessed with Duncan's multiple-range test (DMRT, p < 0.05).

For every treated pot, the dissipation rate (*D*) of Pyr in soils was calculated as $D = (C_0-C_t) \times 1000\%/C_0$, where C_0 was the initial contents of Pyr in soils and C_t denoted residual Pyr.

For given biotic and abiotic factor *i*, its contribution rate (T_i) to the dissipation of PAHs in the process of phytoremediation was expressed as $T_i = D_i \times 1000\%/W \cdot C_0$, where D_i was the removed amount of PAHs by given factor *i* and *W* denoted the weight of spiked soils in pot. Obviously, the *D* value should be theoretically equal to the sum of contribution rates of all factors (Wei and Pan, 2010; Pan *et al.*, 2014).

RESULTS

Plant biomass: As shown in Fig. 1, Sudan grass showed no signs of stress and produced abundant biomass in spiked soils. Throughout the experiment, the



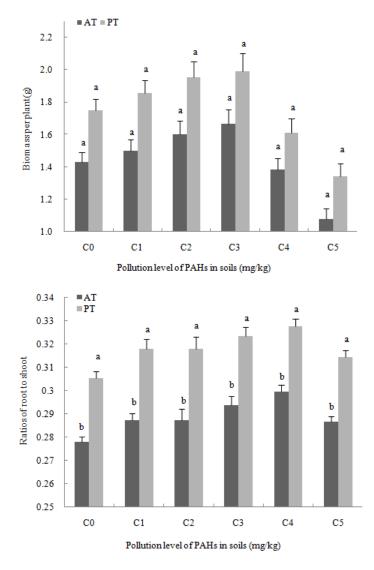


Fig. 1: Biomass and ratios of root to shoot of plants grown in treated soils; PT and AT denoted the Pyr-spiked soils at the presence and absence of 4% Tween 80, respectively

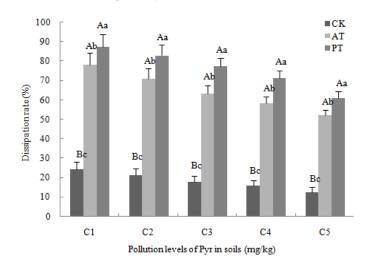


Fig. 2: Dissipation rates of pyrene in soils under different treatments; CK, PT and AT represent unplanted soils and planted soils at the presence and absence of 4% Tween 80, respectively

total mass of Sudan grass in soils with low (C_1) , medium (C_3) and high (C_5) initial contents of Pyr were almost equal to those of unspiked soils (1.43g) with the almost same ratios of root to shoot biomass (0.284). Though biomass of plants grown in high pollution level soils (C5) was slightly decreased, all plants tested did not also show any visible sign of toxicity, indicating that establishment of vegetation in these soils is feasible. The weights of roots or shoots of Sudan grass growing in soils at the presence of 4% Tween 80 (PT) were on average 27.2% and 16.6% greater than those grown in spiked soils, but the differentiation in the total biomass of seedlings growing in variously spiked soils $(C_0 \sim C_5)$ between at the presence of Tween 80 (PT) and absence of that (AT) was insignificant (n = 30, p > 0.05). However, their ratios of root to shoot biomass tended to be significantly greater at the presence of Tween 80 (n = 30, p < 0.05), indicating that the presence of surfactants (Tween 80) was more favorable to the growth of root systems.

Dissipation of HOCs: On the whole, residual Pyr in soils at the presence of 4% Tween 80 (PT) were always lower than those at the absence (AT) for the same treatment. Within the same group, the removal of Pyr in treated soils showed a consistent descending order: CK1STR3SCK2STR4. With an increase of initial contents of Pyr in soils, their D value linearly declined for the same treatment. As shown in Fig. 2, the presence of plants apparently enhanced the removal of Pyr in soils at the presence or absence of Tween 80. During the experiment, approximately 757.33‰ (in a range of 863.94~609.63‰) of Pyr was removed from the TR₄ with surfactants as compared with 665.19%(801.84~539.99‰) degradation without surfactants, while only 182.38‰ (242.28~122.79‰) removal occurred in CK₂ without surfactants at concentrations of $C_1 \sim C_5$. Of the spiked soils, the Pyr dissipation rates in TR₄ with earthworms were the highest, indicating that addition of surfactants to phytoremediation system could enhance removal of Pyr to a certain extent.

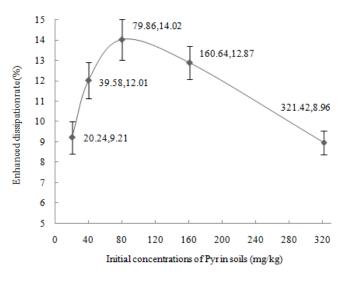


Fig. 3: Enhanced dissipation rates of pyrene in soils at different pollution levels by Tween 80

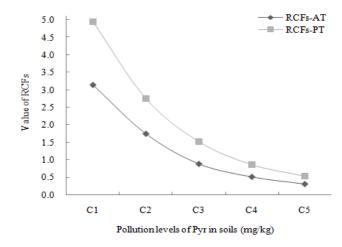


Fig. 4: RCFs of pyrene as a function of PAHs concentrations in soils; RCFs-PT and RCFs-AT denoted root concentration factors in soils at the presence and absence of Tween 80, respectively

Though the dissipation of Pyr in treated soils with surfactants were larger than those without surfactants, the extent of enhanced removal varies greatly among five pollution levels. As shown in Fig. 3, the enhanced dissipation of Pyr in soils with medium pollution level (C_3) was the highest, which were up to 140.16‰ in the vegetated soils (TR₄). Comparatively, fewer extents of removal enhanced occurred in soils with low (C₁) or high (C₅) pollution level, only 92.14‰ and 89.64‰ additional removal, respectively. On a whole, extents of enhanced dissipation of Pyr were up to114.15‰ during the 70d SEPR-experiments.

HOCs accumulated in plant tissues: As expected, accumulation levels of Pyr in plant tissues have a positive correlation with their pollution levels; Pyr contents in plant tissues are always lower in soils at the presence of 4% Tween 80 (PT) than those at the absence (AT). With the increase of pollution levels in soil, Pyr contents in roots and shoots of plants growing in soils without surfactants increased from 14.17 to 72.26 mg/kg and from 2.08 to 13.79 mg/kg while increased from 9.69 to 58.59 mg/kg and from 1.32 to 10.36 mg/kg when adding surfactants, respectively.

Based on residual HOCs in soils and their accumulation levels in roots of plant tissues by the end of the experiment, RCFs (root concentration factors, defined as the ratio of HOCs concentration in roots and one in soils) of Pyr in soils were calculated as shown in Fig. 4. With the increase of contaminants concentration in soils, RCFs of Pyr tended to decrease. Values of RCFs (0.30~3.14) in plant-soil system without surfactants (AT) were always lower than those (0.53~4.95) at the presence of surfactants (PT).

Removal pathways of HOCs: Removal of HOCs in the plant-soil system attributed to both biotic pathways, e.g., plant accumulation, plant metabolism, microbial degradation and plant-microbial interactions and abiotic pathways, e.g., leaching, volatilization, photodegradation and irreversible sorption, etc. If the variation of abiotic loss of HOCs between planted and unplanted soils was negligible, HOCs removed in CK_2 should be equal to the sum of abiotic loss and microbial degradation while loss of HOCs in CK_1 attributed to abiotic loss. Thus, removal of Pyr in soils could approximately be expressed as:

$D_1 =$	T_a
$D_2 =$	$T_a + T_m$

$$D_3 = T_a + T_c + T_d$$

$$D_4 = T_a + T_c + T_d + T_m + T_{pm}$$

where, D_1 , D_2 , D_3 and D_4 were the removed amount of Pyr in CK₁, CK₂, TR₃ and TR₄ and T_a , T_m , T_c , T_d and T_{pm} represented the contribution of abiotic loss, microbial degradation, plant accumulation, plant metabolism and plant-microbial interactions on removal of Pyr, respectively.

As to soils added surfactants, the potential of surfactants in increasing the mass transfer and the bioavailability of these compounds, facilitating their degradation should be also taken into account. Then, enhanced value (ΔD_i^s) of Pyr in CK₁, CK₂, TR₃ and TR₄, could be respectively expressed as:

$$\Delta D_1^{s} = T_a^{s}$$

$$\Delta D_2^{s} = T_a^{s} + T_m^{s}$$

$$\Delta D_3^{s} = T_a^{s} + T_c^{s} + T_d^{s}$$

$$\Delta D_4^{s} = T_a^{s} + T_c^{s} + T_d^{s} + T_m^{s} + T_{pm}^{s}$$

where ΔD_i^s represented the enhanced removed value of Pyr in CK₁, CK₂, TR₃ or TR₄ when adding surfactants and T_a^s , T_m^s , T_c^s , T_d^s and T_{pm}^s did the enhanced amount of T_a , T_m , T_c , T_d and T_{pm} by surfactants, respectively. Based on Pyr detected in leachate, soils, plant and earthworm tissues, contribution rate of each pathway on removal of Pyr was calculated as Table 1.

In the entire SEPR-experiments, no leachate was produced when WHC in soils was maintained at about 45% and abiotic loss by leaching was insignificant. As seen from Table 1, the D value of Pyr in microbeinhibited pots, i.e., CK_1 , was 18.16~32.07‰ (M = 25.58‰), indicating that abiotic loss was a relatively minor pathway for the dissipation of Pyr. However, contribution of microbial degradation on removal of contaminants was notable, which was up to $104.62 \sim 210.23\%$ (M = 158.76‰). In this case, role of plant accumulation in dissipation of Pyr was inappreciable compared to the total loss of HOCs. During the experiment, the amount of Pvr accumulated by Sudan grass only accounted for $0.29 \sim 1.06\%$ (M = 0.61‰) from C_1 to C_5 treatment, which were less than 0.92‰ of Pyr degraded averagely while one derived from plant metabolism was in the range of $5.14 \sim 49.24\%$ (M = 25.09‰), which were less than 38.22‰ of Pyr degraded in soils. Among all removal pathways, the plant-microbial interactions contributed the most part for rhizoremediation of HOCs, which removed up to 456.73‰ (in a range of

Table 1: Contributions of biotic and abiotic factors on PAHs dissipation at the absence or presence of Tween 80 (‰)

Factors	C_1	C_2	C_3	C_4	C ₅
Abiotic loss	32.07 (1.18)	29.28 (1.94)	25.64 (2.28)	22.61 (1.51)	18.16 (1.16)
Microbial degradation	210.23 (15.35)	182.62 (27.82)	152.25 (37.83)	134.34 (34.34)	104.62 (18.54)
Plant accumulation	1.06 (0.00)	0.72 (0.00)	0.62 (0.00)	0.37 (0.00)	0.29 (0.00)
Plant metabolism	49.24 (6.83)	39.46 (5.83)	21.82 (3.95)	9.81 (3.15)	5.14 (2.21)
Plant-microbial interactions	509.24 (39.05)	476.51 (64.32)	452.19 (76.05)	435.83 (69.68)	411.78 (47.69)

Data in bracket were the enhanced T_i of given factor *i* when adding 4% Tween 80

411.78~509.24‰) from spiked soils averagely, illuminating that the predominant pathway responsible for the dissipation of HOCs in soils was the plant-microbial interactions.

As shown in Table 1, enhanced dissipation of Pyr in soils at the presence of 4% Tween 80 varied greatly among removal factors. On the whole, the enhanced extent of plant-microbial interactions and microbial degradation were higher, which was up to 59.36‰ and 26.78‰ on average while comparing with the much lower abiotic loss, whose contributions enhanced only accounted for 1.61‰. It was notable that the contribution rate of plant accumulation to Pyr was slightly lower in soils at the presence of Tween 80 than in soils without surfactants, which was consistent with the experimental facts that with surfactants, Pyr concentrations in plant tissues were always lowered at the same soil concentrations.

DISCUSSION

Phytoremediation is a potential "green", costeffective and secure technology that uses plants for the in-situ cleanup of contaminated sites. It is a promising remediation technique for soils contaminated with organic compounds (Tejeda-Agredano et al., 2013). Plants may assist in the restoration of HOCs through the release of root exudates. Exudates act as microbial substrates that enhance soil microbial populations and their activity in the vicinity of HOC contamination. Moreover, plant uptake and accumulation may also contribute to the removal of HOCs from the soil environment (Sumia et al., 2013). In this study, plantmediated HOCs dissipation was investigated and loss of Pyr in spiked soils $(C_1 \sim C_5)$ growing Sudan grass was 801.84~539.99‰ of these chemicals, which were 2309.22~3397.34‰ larger than those in unplanted soils. The high dissipation ratios of HOCs in planted soils, the evident dissipation promotion of these compounds in the presence of vegetation and the healthy growth of the plant in variously spiked soils suggest the feasibility of remediation of HOCs in soils using Sudan grass.

The addition of surfactants to a soil contaminated with HOCs might increase the mass transfer and the bioavailability of these compounds, facilitating their degradation. The ability of the surfactants to increase desorption, apparent aqueous solubility and microbial bioavailability of HOCs has been well documented. Recently, it has been shown that addition of surfactants facilitates the uptake of HOCs by plants (Gao *et al.*, 2007). In this study, After adding 4% Tween 80, the dissipation of Pyr in variously unplanted soils ($C_1 \sim C_5$) were 257.93~142.43‰, which were 64.53~159.94‰ higher than those in corresponding soils without surfactants and in soils with Sudan grass were up to 863.94~609.63‰, which were 77.27~129.14‰ higher

than those in corresponding soils without surfactants, respectively. These results indicate a promising opportunity to use surfactants to enhance phytoremediation efficacy.

It is noteworthy that the amount of Pyr accumulated by Sudan grass growing in soils added surfactants were always less than those without surfactants and Pyr concentrations in plant tissues, irrespective of root and shoot, were always slightly lower at the same soil concentrations, implying that adding surfactants seemed to be beneficial to decreasing HOCs accumulations in plant tissues to a certain extent, which could be especially beneficial for relieving potentially ecological risks.

CONCLUSION

In conclusion, adding surfactants apparently reinforced the dissipation of Pyr in planted soils at initial contents ranging from 20.24 to 321.42 mg/kg. During the 70-d SEPR-experiments, after adding surfactants, loss of Pyr in variously spiked soils ($C_1 \sim C_5$) with Sudan grass were up to 863.94~609.63‰, which were 92.06‰ larger than those in corresponding soils without surfactants averagely, which suggested that SEPR may be a feasible way for reinforcing removal of HOCs in contaminated sites. However, it remains to be seen whether the dissipation of HOCs in the plant-soil system adding surfactants is as effective under large-scale field conditions as it appears under laboratory conditions.

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